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Abstract

Benthic macro – invertebrates (BMI) are good indicators of surface water quality but aquatic pollution in recent times threatens their existence. Information on anthropogenic effluents, BMI abundance and diversity in Gbalegbe River is limited, hence this study. Gbalegbe River (12.5 Km) was spatially divided into eight stations (S1, S2, S3, S4, S5, S6, S7 and S8) comprising three sampling units per station based on dominant anthropogenic activities. Water and BMI samples were collected from each station monthly for 24 months following standard methods. Water samples were analysed for Dissolved Oxygen (DO, mg/L) and Temperature (°C) using standard procedures. The BMI samples collected were identified to species level and counted. Species abundance and diversity (Simpson and Shannon) indices were calculated. Data were analysed using descriptive statistics and ANOVA at $\alpha_{0.05}$. The highest (4.97 ± 0.36) and least (3.12 ± 0.17) DO levels were obtained in S1 and S2, while water temperature ranged from $25.68 \pm 3.34^\circ\text{C}$ to $27.91 \pm 0.03^\circ\text{C}$ in S3 and S2. Dry and wet seasons recorded 5.79 ± 0.43 and 3.57 ± 0.59 as highest and least DO, while water temperature ranged from $27.85^\circ\text{C} \pm 0.94$ to $27.99^\circ\text{C} \pm 5.06$ in dry and wet seasons. A total of 1186 individual BMI belonging to 6 Orders, 15 Families and 16 Species were recorded. Stations 1 and 2 recorded the highest 236 (19.9 %) and least 64 (5.4 %) abundance of BMI. Wet and dry seasons recorded *Ephemera doris* 55 (8.0 %) and *Tricorythodes albilineatus* 10 (2.0 %) as highest and least abundant BMI. Highest (0.99) and least (0.86) values of Simpson and Shannon (2.05, 2.98) occurred in S2 and S1. Seasonally, Simpson (0.95, 0.92) and Shannon (2.74, 2.69) were higher in wet season than in dry season. Due to the relatively low levels of DO and Shannon diversity index obtained, BMI abundance of Gbalegbe River could be negatively affected.

Keywords:

Surface water, Species diversity, anthropogenic effluents, Aquatic pollution.

Introduction

Gbalegbe River is the main river flowing through Ughelli Town, Delta State. The river serves as a means of transportation, fishing, water for agricultural activities and domestic purposes (Ewutanure and Olaifa, 2017). According to Ewutanure and Olaifa, (2018), Ughelli Town was initially known for farming activities but has been highly urbanized leading to the location Power Holding Company of Nigeria (PHCN) Nigeria Limited, oil (SHELL, Nigeria National Petroleum Company) and construction (Setraco Nigeria Limited, Julius Berger Nigeria Limited), Glass (Beta glass Nigeria Limited), Rubber, Battery repairer, Sand mining and Cassava companies. In 1958, petroleum resources were detected within the town, but oil mining started in 1965. Since then, crude oil from the Ughelli fields has been transported through the 225 Km Trans – Niger Pipeline South – Eastward to the Port of Bonny for export (Ochukoet al., 2008).

The location was chosen because it is the major area where PHCN, Petroleum, Glass, Rubber production, Gas flaring, Sand mining activities and Cassava were taking place in Delta State (Ewutanure and Olaifa, 2017). Anthropogenic effluents from these companies are directly or indirectly discharged untreated through drainages into Gbalegbe River thereby polluting the breeding ground of both benthic macroinvertebrates and fishes (Ewutanure and Olaifa, 2018a). The cumulative effects of untreated effluents in the aquatic ecosystem have the potential of altering its physical, chemical and biological qualities (Salman et al. 2011).

Benthic macroinvertebrates are good indicators of surface water quality because they are affected by the physical, chemical and biological constituents of water bodies (Popoola and Otalekor, 2011). They show the impacts of acute, chronic and accumulative toxic effects of polluted environment (Bouldot,

2014). Rivers are vital aspects of the hydrosphere that provide habitats for diverse flora and fauna species (Salman et al. 2011). But, in recent times, they are being subjected to numerous anthropogenic activities capable of altering and modifying the aquatic ecosystem leading to a deviation from the natural ecological composition all over the globe (Ewutanure et al. 2022a).

Anthropogenic activities could bring about species and habitat loss, habitat fragmentation and a shift in community structure (Ewutanure and Binyotubo, 2021). Due to the relevance of the aquatic ecosystem to human survival, the significance of its biodiversity and the awareness that the aquatic ecosystem serve as the primary sink for pollutants of anthropogenic and natural origin becomes of major concern (Xiaodong et al. 2010).

Benthic macroinvertebrates inhabit the sediments of all aquatic ecosystem such as rivers, ponds, swamps, estuary, lakes, streams and oceans (Popoola and Otalekor, 2011; Lennie and Dean, 2013). The common groups of benthic macroinvertebrates are worms (polychaetes and oligochaetes), molluscs (bivalves), gastropods and crustaceans (amphipods and decapods) (Echols et al. 2010).

Based on the position that benthic macroinvertebrates occupy in the benthic ecosystem, they can be differentiated into: Infauna benthic macroinvertebrates that live in sediments by burrowing, almost all worms and bivalves belong to this category, and Epifauna which are benthic macroinvertebrates that live on the surface of bottom sediments. Many crabs and gastropods are known to be epifauna (Shama and Rawat, 2009).

Some groups of the epifauna attached themselves to hard surfaces of the banks or pilings within water bodies (Xiaodong et al. 2010). Benthic macroinvertebrates play essential role in

energy transfer through the ecosystems by filtering plankton and also acting as a food source to fish thereby linking primary production with higher trophic levels (Fengqing *et al.* 2012).

Bouldot (2014) reported that benthic organisms range from bacteria to plants (phytobenthos) and animals (zoobenthos) and are found within the different levels of the food web. However, with respect to sizes, benthic invertebrates can be classified into: Microbenthos, <0.063 mm; Meiobenthos, 0.063–1.0 mm; Macro-benthos >1.0 mm and Megabenthos > 10.0 mm.

Benthic macroinvertebrates degrade organic matter before bacteria remineralization, rework and oxygenate the bottom sediments through their activities (Friberget *et al.* 2010). They serve as biological indicators because they are capable of giving information on the health integrity of the aquatic ecosystem through the sensitivity of indicator species or through some features which might trigger environmental stress over a period of time (Amore *et al.* 2009).

The survival of benthic macroinvertebrates could be threatened when exposed to chemical contaminants in sediment and hypoxia or anoxia situations as a result of organic materials break down, limited mobility that restricts their escape capability, taxonomic and functional diversity that make them suitable for detecting different types and levels of pollutants (Arimoro and Ikomi, 2008).

Previous studies on the use of benthic macroinvertebrates' community in assessing the health status and water quality of rivers in the tropics were limited elaborate (Oku *et al.* 2014; Popoola and Otalekor, 2011; Arimoro and Ikomi, 2008; Deliz – Quinones, 2005). Hence, this study aims at evaluating the benthic macroinvertebrate community of Gbalegbe River, Delta State, Nigeria.

Materials and methods

Location of Study Area

Gbalegbe River with total length of 12.5 Km is located within latitudes 05°10'N and 05°17'N of the Equator and Longitudes 005°56'E and 005°13'E of the Greenwich meridian. It has its source from the Asaba–Ase River, Delta State (Ewutanure and Olaifa, 2018a).

Climate and vegetation of the study area

The wet (March – October) and dry (November – February) seasons (Ewutanure and Olaifa, 2017) are the two climatic seasons prevalent in the study area. The wet season is characterized by heavy rainfall, while occasional downpour is experienced during the dry season (Ewutanure and Olaifa, 2017). The study area has a mean yearly rainfall of 2700mm, while the rainfall peaks in June/July and September with a mean annual temperature of 27 °C (Ewutanure and Olaifa, 2021a). According to Ewutanure and Olaifa, (2021b) the vegetation of the study area is rain forest with adjoining low-lying land that is permanently waterlogged.

Sampling Stations

Gbalegbe River was spatially stratified into eight stations (S1 – low human activities; S2 – glass production factory; S3 – power plant; S4 – rubber processing mill; S5 – Oil farm tanks; S6 – Battery repairer shops; S7 – Cassava processing mill and S8 – Sand mining) in relation to closeness to major anthropogenic activities (Ewutanure and Olaifa, 2018a). In each station, three sampling points were randomly selected. Temporal stratification covered wet (March – October) and dry (November – February)

(Ochuko *et al.* 2008) seasons. Sampling was done monthly for 24 months.

Collection of water samples

Water samples from the surface (0 – 15cm) were collected using polyethylene bottles after sterilization with nitric acid and rinsing with the water from the sampling stations 3 – 4 times (AOAC, 1990; ASTM, 2006). Surface water samples were collected by lowering the polyethylene bottle just below the surface, opening and closing it by hands before taking it out (Ewutanure and Olaifa, 2018a).

Sampling Techniques

Determination of Physico-chemical parameters of water

Water Temperature was measured using mercury in glass thermometer, pH with digital pH meter (Hanna model: HI – 98107, USA) and dissolved oxygen was determined by Winkler's method (Gupta, 2001). Samples for biochemical oxygen demand (BOD₅) was determined as described by Boyd, (1979). Turbidity was measured by using a turbidometer in accordance with USEPA, (AOAC, 1990) standards.

Collection, preservation and identification of benthic macro – invertebrates

A modified van Veen bottom grab (van Veen, 1933) sampler was used for sediment collection for the identification of benthic macroinvertebrates. Since most of the benthic macroinvertebrates inhabit the top 5 – 10 cm below the river bed, the van Veen bottom grab option was considered suitable because it can dig up to 5 – 10 cm in depth on river beds (Eleftheriou and Holme, 1984).

Collected sediments were emptied into a stainless-steel bucket containing water (Popoola and Otalekor, 2011). Thereafter, the sediments were filtered thoroughly through a sieve of mesh size of 0.5 mm and the organisms with the debris stored in a wide-mouthed container and preserved in 10% formalin solution. In the laboratory, collected benthic macroinvertebrates were sorted from the debris and identified with the aid of a compound microscope (X 100), using relevant taxonomic keys and pictures (Macan, 1999; Heckman, 2002; Lynne, 2004) to species level.

Statistical analysis

Data obtained from this study were subjected to descriptive and inferential statistics by using SPSS (version, 20). Microsoft Excel, (version 2010) was used to calculate species abundance. Data were pooled and presented as spatial and temporal mean variances to evaluate if their differences were significant at $\alpha_{0.05}$, while Shannon–Wiener (Pielou, 1969), Simpson (Simpson, 1949), Richness, Evenness and Margalef (Margalef, 1969) indices were calculated by using Paeleontological Statistics (Past), version 3.16.

Results and Discussion

Physico – chemical parameters recorded among stations and between seasons are presented in Tables 1 and 2. Compared with Station 1, significant differences ($P > 0.05$) were visible. Spatially, DO concentration was highest in Stations 1 and least in Station 2. Levels of Temperature, BOD and Turbidity were highest in Station 2, while their least concentrations were recorded in Stations 3 and 1, respectively.

Concentration of dissolved oxygen obtained throughout this study ranged from 3.12 – 5.79 mg/L. These values were within the recommended ranges for warm water fish (Boyd, 1979). This

report agrees with earlier reports on Okerenkoko Estuary (Ewutanure and Olaifa, 2021a); Gbalegbe River Delta State, Nigeria (Ewutanure and Olaifa, 2017); Sagbama Creek, Niger Delta (Seiyabohet *et al.* 2017); Nun River (Ogambaet *et al.* 2015); River Areba at Olomoro, Delta State (Idodo – Umeh, 2003). The slight fluctuation in the level of DO observe could be associated with the untreated anthropogenic effluents from PHCN, Oil (SHELL, Nigeria National Petroleum Company), construction (Setraco Nigeria Limited, Julius Berger Nigeria Limited), Glass (Beta glass Nigeria Limited), Rubber, Battery repairer, Sand mining and Cassava companies(Ewutanure and Olaifa, 2018a).

Mean water temperature recorded during the sampling period ranged from 25.77°C to 27.91°C. This is in consonance with the temperature range (25 – 30 °C) for tropical water fisheries (Boyd, 1979; FEPA, 1991). The report also corroborates the observations from Okerenkoko Estuary (Ewutanureet *et al.* 2022b); Gbalegbe River (Ewutanure and Olaifa, 2018a); rivers and creeks in the Niger Delta region (Kwenet *et al.* 2012); Igbedi Creek (Seiyabohet *et al.* 2013) and Epie Creek in Bayelsa State (Seiyabohet *et al.* 2016).

Concentrations of BOD recorded during the study were within the recommended range of 5.0 mg/L of BOD (Boyd, 1979; FEPA, 1991), but significant differences were observed among stations. Elevation in BOD could be associated with increased anthropogenic activities in Gbalegbe River (Ram *et al.* 2011; Ewutanureet *et al.* 2022b). An increased level of BOD could cause a decrease in the concentration of DO (Ewutanure and Olaifa, 2021b). A sustained and prolonged elevation in the value of BOD could lead to an anoxic condition in an aquatic ecosystem (Ewutanure and Olaifa, 2021c).

The levels of pH recorded during this study were within the established pH range of 6.6 – 9.0 (FEPA, 1991) and 6.5 – 8.0 (Boyd, 1979). Variation in pH levels recorded was in agreement with report on Gbalegbe River, Delta State (Ewutanure and Olaifa, 2018b); New Calabar River Ekeh and Upper Nun River of the Niger Delta Region of Nigeria (Kwen *et al.* 2012).

Wet season recorded higher concentration of turbidity than in dry season. Higher turbidity values than10 FTU as recommended by Boyd, (1979) and FEPA, (1991) could be attributed to the increased presence of suspended solids, organic and inorganic compounds present in the River (Ewutanure and Olaifa, 2021d). Seasonally, increased concentrations of

Temperature, BOD and Turbidity recorded in the wet season than in the Dry Season could be attributed to increased rate of sand mining activities and heavy rainfall (Ogagaet *al.* 2015; Ayandiranaet *al.* 2018).

The compositions, distribution and abundance of benthic macroinvertebrate species among stations and between seasons are presented in Tables 3 and 4, respectively. A total of individual numbers of 1186 benthic macroinvertebrates belonging to 6 orders, 15 families and 16 species were recorded throughout the study period. Spatially, Station 1 recorded the highest abundance of benthic macro – invertebrates 236 (19.9 %). Stations 8 and 7 ranked next with the abundance of benthic macro – invertebrates sampled as 206 (17.4 %) and 203 (17.1 %), while the least 64 (5.4 %) occurred in Station 2. Highest 409 (34.5 %) and least 53 (4.5 %) abundant benthic macro – invertebrates recorded among stations were *Ephemeroptera* and *Coleoptera*.

Seasonally, highest 689 (58.1 %) and least 497 (41.9 %) individual number of benthic macro – invertebrateswere recorded in wet and dry seasons, respectively. Ephemeroptera 245 (35.6 %); 169, (34.0 %) occurred as highest orders in wet and dry seasons, while the least orders were Coleoptera 33 (4.8 %) and Tricoptera 20 (4.0 %), respectively. Tricorythidae 75 (10.9 %) and Libellulidae 20 (2.9 %) occurred as most and least abundant families of macro – benthic invertebrates in the wet season, but Corixidae and Ephemerellidae 45 (9.15 %) occurred as most abundant families, while the family Culicidae 10 (2.0 %) was recorded as the least abundant.

Wet season recorded *Ephemerelladoris* 55 (8.0 %) as most abundant macro – benthic invertebrates, *Hesperocorixacastanea* 54 (7.8 %) ranked next, while the least abundant was *Pantataflarescens* 20 (2.9 %). The most 44 (8.9 %) and least 10 (2.0 %) abundance benthic macro – invertebrates recorded in the dry season were *Tricorythodesalbilineatus* and *Aedesaegypti*. Conductive water quality parameters determine the spatial and seasonal distribution of benthic macro – invertebrates’ compositions and abundance in any aquatic water environment (Davies and Ekperusi, 2021; Seiyabohet *et al.* 2017).

Table 1. Mean Physico – chemical parameters of water from Gbalegbe River

Parameter	Stations							
	S 1	S 2	S 3	S 4	S 5	S 6	S 7	S 8
DO (mg/L)	4.97±0.36 ^a	3.12±0.17 ^b	3.75±0.71 ^a	3.89±0.15 ^a	3.41±0.35 ^a	3.98±0.24 ^a	3.44±0.70 ^a	3.81±0.39 ^a
Temp(°C)	27.02±1.2 ^{5b}	27.91±0.03 ^a	25.68±3.3 ^{4c}	26.40±0.9 ^{8b}	26.95±0.1 ^{5b}	27.23±0.0 ^{0b}	27.37±0.2 ^{4b}	25.77±0.1 ^{5c}
BOD (mg/L)	0.48±0.19 ^b	1.46±0.13 ^a	1.06±0.16 ^a	1.14±0.01 ^a	1.29±0.09 ^a	1.03±0.19 ^a	0.87±0.14 ^c	1.40±0.11 ^a
pH	7.17±0.41 ^a	6.25±0.79 ^a	7.67±0.16 ^a	7.30±0.32 ^a	7.79±0.37 ^a	7.50±0.14 ^a	7.58±0.10 ^a	7.11±1.05 ^a
Turbidity (FTU)	7.81±1.87 ^c	45.49±10.2 ^{1a}	32.13±7.5 ^{4c}	37.15±8.0 ^{9b}	36.01±4.5 ^{4b}	25.19±3.5 ^{4c}	40.02±7.4 ^{2b}	35.54±9.7 ^{9b}

Note: S 1 – S 8 = Station 1 to Station 8. DO = Dissolved oxygen, Temp = Temperature, BOD = Biological oxygen demand

Table 2. Mean Physico – chemical parameters of Gbalege River monitored in wet and dry seasons

Parameters	Seasons		Boyd, (1979)	FEPA, (1991)
	Wet	Dry		
DO (mg/L)	3.57±0.59 ^b	5.79±0.43 ^a	5 – 10	> 5
Temp. (°C)	27.99±5.06 ^a	27.85±0.94 ^a	25 – 32	20 – 33
BOD (mg/L)	1.98±0.32 ^a	1.56±0.07 ^a	5.0	5.0
pH	7.61±0.16 ^a	7.98±1.35 ^a	6.5 – 8	6.5 – 9
Turbidity (FTU)	35.35±0.91 ^b	31.17±2.34 ^a	10.0	10.0

Note: DO = Dissolved oxygen, Temp. = Temperature, BOD = Biological oxygen demand

Table 3. Compositions, distribution and abundance of benthic macro – invertebrates of Gbalege River among stations

Orders	Families	Genus and Species	Numbers of benthic macro – invertebrates collected per station								Total	% Abundance
			1	2	3	4	5	6	7	8		
Hemiptera	Corixidae	<i>Hesperocorixacastanea</i>	0	7	21	11	1	1	1	57	99	8.3
Hemiptera	Gerridae	<i>Gerrisremiges</i>	0	8	1	4	23	2	2	21	61	5.1
Hemiptera	Cicadellidae	<i>Lonaturamegalopa</i>	0	8	3	12	34	8	3	18	86	7.3
		Sub-total	0	23	25	27	58	11	6	96	246	20.7
Ephemeroptera	Epemerellidae	<i>Ephemerelladoris</i>	30	0	2	9	2	11	23	23	100	8.4
Ephemeroptera	Ephemeridae	<i>Hexagenialimbata</i>	11	0	2	3	1	9	7	19	52	4.4
Ephemeroptera	Heptageniidae	<i>Stenonemaexiguum</i>	14	0	1	2	9	7	5	4	42	3.5
Ephemeroptera	Isonychiidae	<i>Isonychiaarida</i>	19	0	1	2	1	3	10	7	43	3.6
Ephemeroptera	Polymitarcyidae	<i>Tortopusincertus</i>	15	0	1	14	1	4	13	5	53	4.5
Ephemeroptera	Tricorythidae	<i>Tricorythodesalbilineatus</i>	19	0	3	2	3	4	76	12	119	10.0
		Sub-total	108	0	10	32	17	38	134	70	409	34.5
Trichoptera	Hydropsychidae	<i>Hydropsychid</i> sp.	10	0	2	8	12	3	9	1	45	3.8
		Sub-total	10	0	2	8	12	3	9	1	45	3.8
Odonata	Aeshnidae	<i>Helocorduliaselysii</i>	0	9	13	7	16	2	3	4	54	4.6
Odonata	Libellulidae	<i>Pantataflarescens</i>	0	5	16	11	12	3	9	7	63	5.3
Odonata	Aeshnidae	<i>Ashnainterrupta</i>	0	4	7	6	9	5	1	8	40	3.4
		Sub-total	0	18	36	24	37	10	13	19	157	13.2
Diptera	Culicidae	<i>Aedesaeegypti</i>	7	1	4	3	6	5	8	3	37	3.1
Diptera	Culicidae	<i>Culexpiens</i>	3	5	2	2	3	7	12	4	38	3.2
		Sub-total	10	6	6	5	9	12	20	7	75	6.3
Coleoptera	Gyrinidae	<i>Gyrinus narrator</i>	1	7	2	20	13	3	2	5	53	4.5
		Sub-total	1	7	2	20	13	3	2	5	53	4.5
	Others	Fish larvae	54	0	3	4	7	2	8	5	83	7.0
		Fish eggs	21	2	5	2	0	0	4	2	36	3.0
		Worms	32	8	7	13	9	5	7	1	82	6.9
		Sub-total	107	10	15	19	16	7	19	8	201	16.9
		Total	236	64	96	135	162	84	203	206		
		%Abundance	19.9	5.4	8.1	11.4	13.7	7.1	17.1	17.4		

Table 4. Compositions, distribution and abundance of benthic macro – invertebrates of Gbalegbe River wet and dry seasons

Order	Families	Genus/species	Wet season total	% Wet season	Dry season total	% Dry season
Hemiptera	Corixidae	<i>Hesperocorixa castanea</i>	54	7.8	45	9.1
Hemiptera	Gerridae	<i>Gerris remigis</i>	28	4.1	33	6.6
Hemiptera	Cicadellidae	<i>Lonatura megalopa</i>	47	6.8	40	8.0
		Sub-total	129	18.7	118	23.7
Ephemeroptera	Epemerellidae	<i>Ephemerella doris</i>	55	8.0	45	9.1
Ephemeroptera	Ephemeridae	<i>Hexagenia limbata</i>	37	5.4	22	4.4
Ephemeroptera	Heptageniidae	<i>Stenonema exiguum</i>	28	4.1	14	2.8
Ephemeroptera	Isonychiidae	<i>Isonychia arida</i>	25	3.6	26	5.2
Ephemeroptera	Polymitarcyidae	<i>Tortopus incertus</i>	25	3.6	18	3.6
Ephemeroptera	Tricorythidae	<i>Tricorythodes albilineatus</i>	75	10.9	44	8.9
		Sub-total	245	35.6	169	34.0
Trichoptera	Hydropsychidae	<i>Hydropsychids</i> sp.	39	5.7	20	4.0
		Sub-total	39	5.7	20	4.0
Odonata	Aeshnidae	<i>Helocordulia selysii</i>	31	4.5	23	4.6
Odonata	Libellulidae	<i>Pantata flarescens</i>	20	2.9	43	8.7
Odonata	Aeshnidae	<i>Ashna interrupta</i>	24	3.5	16	3.2
		Sub-total	75	10.9	82	16.5
Diptera	Culicidae	<i>Aedes aegypti</i>	27	3.9	10	2.0
Diptera	Culicidae	<i>Culex pipiens</i>	26	3.8	12	2.4
		Sub-total	53	7.7	22	4.4
Coleoptera	Gyrinidae	<i>Gyrinus narrator</i>	33	4.8	20	4.0
		Sub-total	33	4.8	20	4.0
	Others	Fish larvae	40	5.8	23	4.6
		Fish eggs	23	3.3	13	2.6
		Worms	52	7.5	30	6.0
		Sub-total	115	16.7	66	13.3
		Total	689		497	
		% Abundance	58.1		41.9	

Tables 5 and 6 show the spatial and seasonal benthic macroinvertebrate diversity index of Gbalegbe River. Spatial trend of Dominance index for Gbalegbe River ranged from 0.01 (S 1) to 0.14 (S 2); Simpson (0.86, 0.99); Shannon (2.05, 2.98) in S 2 and S 1; Evenness (0.10, 0.75) in S 8 and S 3 and Margalef (2.02, 3.51) in S 2 and S 1. Seasonally, Simpson, Shannon and Evenness were higher in wet season than in dry season, while Dominance and Margalef values were higher in dry season than in wet season, respectively. Benthic macro – invertebrates’ species diversity is expressed as the number or combined abundance of the species (Usman *et al.* 2014).

Table 5. Diversity index of benthic macro – invertebrates of Gbalegbe River among stations

Index	Stations							
	S 1	S 2	S 3	S 4	S 5	S 6	S 7	S 8
Individuals	236	64	96	135	162	84	203	206
Dominance (D)	0.01	0.14	0.06	0.07	0.08	0.07	0.05	0.03
Simpson (1 – D)	0.99	0.86	0.94	0.93	0.92	0.93	0.95	0.97
Shannon (H)	2.98	2.05	2.10	2.74	2.33	2.49	2.52	2.70
Evenness (E)	0.42	0.54	0.75	0.54	0.41	0.51	0.48	0.10
Margalef	3.51	2.02	2.53	2.42	2.25	2.22	2.53	2.12

Table 6. Diversity index of benthic macro – invertebrates of Gbalegbe River between seasons

	Wet Season	Dry Season
Individuals	689	497
Dominance (D)	0.05	0.08
Simpson (1 – D)	0.95	0.92
Shannon (H)	2.74	2.69
Evenness (E)	0.42	0.54
Margalef	2.11	3.09

Maximum value of Dominance obtained in S 2 depicted complete dominance of few species (Chessman, 2009). In biological communities, Shannon – Wiener indices ranged from 0 to 5 (Shannon – Wiener, 1949). With regards these indices, values less than 1 represent heavily polluted state, values ranging from 1 to 2 are equivalent to areas of moderate pollution, while values greater than 3 depicts stable and healthy environmental state (Durance and Ormerod, 2009).

It has been reported that, Simpson index ranges from 0 to 1 (Ewutanure and Binyotubo, 2021). It has been established that in relatively undisturbed aquatic ecosystem, diversity index ranges from 0.6 to 0.9, while surface water impacted by anthropogenic effluents shows very low diversity which could be near zero value (Boudot, 2014). It has been reported that Simpson diversity index is always higher where the community is dominated by less number of species and when the dominance is shared by large number of species (Popoola and Otalekor, 2011). Whenever Simpson diversity index increases towards higher values, the evenness index goes in a reversed order and vice versa (Xiaodong *et al.* 2010). Hence, it is believed that Simpson and Evenness possess an inverse relationship (Lennie and Dean, 2013; Suleman *et al.* 2015).

The low diversity obtained in S 2, as described by the Shannon and Margalef index, may be due to fewer numbers of species and persistent environmental perturbation due to increased anthropogenic activities (Salman *et al.* 2011). Margalef is highly sensitive towards sample size (Echols *et al.* 2010). Evenness is the inverse of the proportion of individuals in the community that belong to the single most common species (Chessman, 2009). Hence, it is determined by its sensitivity to sample size as Margalef's index (CASQA, 2014). Like Shannon index, this index does not provide any information regarding the rare species in the population (Ewutanure *et al.* 2022b).

It is an established fact that Margalef index has no specified value as a limit but it could show some fluctuation in values depending on the number of species, hence, it could be applied in comparison among stations or study sites (Shama and Rawat, 2009). Margalef index only considers species richness which reflects species sensitivity to sample sizes (Margalef, 1968; Fengqing *et al.* 2012). Margalef has a value greater than 1 unlike the Simpson index whose values range from 0 to 1 and it could compare favourably with species richness of various study area over Simpson index (Ewutanure and Olaiifa, 2018a).

Conclusions

Generally, species composition, distribution and abundance for benthic macro – invertebrates of Gbalegbe River showed significant variations among stations and between wet and dry seasons, while the information provided by majority of the diversity indices could be used to determine (i) the habitat qualities of benthic macroinvertebrate communities of a river (ii) aquatic ecosystem component with regard to Evenness and their interactions with benthic macro invertebrate communities that are predominant in an area. Diversity indices could indicate high values for a relatively undisturbed aquatic ecosystem, while degraded ones exhibit low values in such a way that environmental quality could threaten the existence of macroinvertebrate communities in river.

It was observed that as Shannon – Weiner and Simpson index increase, evenness also increases but with varying trends. It is therefore recommended that regular and consistent bio – monitoring of Gbalegbe River should be done to ensure the sustenance and species abundance stability of its benthic macroinvertebrates communities.

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